

Effects of polymer dosage on alum sludge dewatering characteristics and physical properties

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Abstract

The proper use of polymers as conditioners is a critical aspect of dewatering processes. In this study, we investigate their physical properties, i.e. size, density and fractal dimension and correlate them with their dewatering characteristics (bound water content, CST and SRF) on alum sludge with cationic polymer conditioning. Using CST measurement to determine the optimum polymer dose may lead to an over-dosing for polymer conditioned sludge. Bound water depletion and interstitial water formation significantly affect the moisture content during polymer conditioning. Moreover, the magnitude of bound water content reflects net change in moisture content, which is decreased by bound water depletion and increased by water formation. Floc size and density measurement suggest poor dewatering performance and increased bound water content are attributable to enlargements in floc size and decreases in floc density. Experimental results indicate that increases in bound water and decreases in floc density are caused by variations of both floc size and aggregation configuration type, not degree of floc compactness.

Keywords: Alum sludge; Bound water; Dewatering; Dilatometer; Fractal dimension; Polymer; Sludge conditioning

1. Introduction

Sludge dewatering aims to reduce the cost of sludge handling and transportation. Synthetic polymer conditioning is an efficient process most frequently used to improve sludge dewatering. Proper use of polymers as conditioners is a critical aspect of sludge dewatering. Therefore, a more thorough understanding of the change in bound water content and the physical properties of sludge in a dewatering mechanism may provide valuable information regarding the optimal use of polymer conditioners in sludge dewatering treatment. The water in sludge plays an influential role in defining

the dewatering characteristics, and is generally classified as bound water and free water. Vesilind [1] has suggested that the bound water content of sludge is a gross estimate of water bound in the sludge in different ways, including as interstitial water, vicinal water, and hydration water. Such a gross estimate can be regarded as the theoretical limit of mechanical dewatering. Several comprehensive studies have measured bound water in sludge [2]. Previous methods, although widely employed in estimating the bound water content of sludge, are limited to operational definitions with respect to measurement methods [3,4]. Among those methods, dilatometry consumes the least amount of time and is easy to operate. This method is based upon the assumption that the bound water remains unfrozen at temperatures

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below free water's freezing point [5]. Moreover, the amounts of total and free water can be determined, thus allowing the sludge's bound water content to be calculated. Smith and Vesilind [6] have suggested that the freezing temperature, the solid concentration and air liberated from the sludge all affect frozen water measurement.

Floc size, density and fractal dimension characterize the sludge's physical properties. Fractal dimension analysis, which measures how particles fill spaces, is highly effective in investigating floc structure [7]. A previous study has surveyed available techniques for analyzing the fractal aggregates [8]. Sato and co-workers [9] have demonstrated that a correlation can be found between bound water and cake solid concentrations in conditioned alum sludge. Smollen [10] has found a similar correlation in biological sludge. Katsiris [11] has reported that the bound water content of sludge decreased with an increase in polymer dosage and suggested that the decrease is due to polymer-water exchange at the binding sites of the sludge particles. Robinson et al. have proposed a similar hypothesis [3]. However, these investigations neglected the effects of variations in physical properties of sludges as a function of polymer dosage.

Sludge microproperties such as floc size and density play influential roles in defining both the rate and extents of waste sludge dewatering mechanisms. Dulin et al. [12] have shown that incorporating dissolved organic carbon into the floc structure may result in a significant decrease in floc density and a correspondingly marked decrease in dewatered solid concentrations. Moreover, Knocke et al. [13] have employed a water balance calculation to verify that the polymer conditioning enhanced the release of floc water, as evidenced by an increase in floc density.

Despite the numerous studies dealing with bound water measurement, the relationship between the sludge's physical properties and dewatering characteristics remains unclear. In this study, we examine the effects of cationic polymer dosage on the physical properties of alum sludge and dewatering characteristics. Specific changes in floc density and floc size on bound water during polymer conditioning are also discussed.

2. Experimental methods and materials

2.1. Sludge sample and polymer

Sludge was obtained from the settling basin at the Chung-Hsien water treatment plant (Taipei, Taiwan). The suspended matter in the surface water was coagulated by adding aluminium sulfate (alum). Table 1 presents the typical sludge characteristics.

Cationic polyelectrolyte (polymer PC-320) was obtained from the Taiwan Polymer Company. Polymer PC-320 is a copolymer of acrylamide and diallyldimethyl-ammonium chloride, with an average molecular weight of $1.1\text{--}1.2 \times 10^7$, and a 20% charge density. Polymer solution (0.1% w/w) was prepared according to methods proposed by the polymer manufacturer.

2.2. Sludge conditioning and dewatering

The mixing apparatus used for sludge conditioning consisted of a 1 l mixing chamber, a stirrer, and a paddle impeller. 800 ml sludge sample was poured into the mixing chamber and mixed with polymer at a paddle rotation speed of 125 rpm for 60 s. After conditioning, 300 ml of the sludge was immediately withdrawn for measuring dewaterability, floc density and further image analysis. The remaining sludge was settled for 2 h, after which the supernatant was discarded. Next, bound water content and dry solids content of the settled sludge were measured. A centrifugation process was also used for sludge dewatering. Settled sludge was centrifuged at 3000 rpm for 30 min. After centrifugation, the supernatant was decanted and the dry solids content of sludge cake was measured after drying at 105°C for 24 h. All experiments were conducted at 20°C.

2.3. Dewaterability measurement

Sludge dewaterability were evaluated by the capillary suction time (CST) and the specific resistance of filtration (SRF). Triton CST apparatus model 200 with a 1.8 cm diameter cylinder and Whatman No. 17 filter paper were used to measure CST value. A standard Buchner funnel apparatus

Table 1
Typical characteristics of alum sludge in the Chung-Hsien water treatment plant

pH	Solid content (%)	Dry density (kg m ⁻³)	CST ^a (s)	Viscosity (cps)	Zeta potential (mV)
7.3	3.82	2278	82	46.8	–12.2

^aCapillary suction time.

with a 9 cm funnel was used for the SRF determinations. A 100 ml sludge sample was filtered with 15 cm · Hg pressure. The filtrate quantity was collected as a function of time.

2.4. Sludge floc density analysis

A free settling test was employed to measure the wet density of sludge flocs. The settling travel of individual sludge aggregates in a quiescent column was recorded by a video camera equipped with a close-up lens. The column is an acrylic glass cylinder (10 cm diameter and 60 cm high) with a glass plate facing the camera. Sludge supernatant was served as the settling medium. Sludge flocs were carefully collected by a pipette and then slowly released into the settling column. The floc diameter and terminal velocity in the column were determined by replaying the tape. Using the measurement of floc terminal settling velocity and floc diameter, floc densities were calculated according to a modified Stoke's equation [14] and presented as wet density. Finally, the dry particle density was measured with a pycnometer.

2.5. Bound water content measurement

Dilatometer and an expression test were employed to measure the sludge's bound water content. The total volume of the dilatometer was 65 ml. A 15 g sludge sample was introduced into the dilatometer and then the rest of the volume was filled with indicator fluid. A mineral oil (Shell Donax TG, USA) was selected as the indicator fluid according to the procedures described by Robinson and Knocke [3]. Dry ice in an ethanol bath was used to reduce the sample's temperature from 20 to –20°C. The initial and final liquid levels on the dilatometer during cooling period were then recorded. The total water content of

sludge was determined by drying at 105°C for 24 h; the frozen water content were determined according to Eq. (1):

$$\text{Frozen water content} = (\Delta L + W \times A)/B \quad (1)$$

where ΔL is the level difference of 20 to –20°C, W is the weight of oil used, A is the oil contraction coefficient for each dilatometer, and B the expansion coefficient of sludge filtrate for each dilatometer. The amount of bound water was calculated by subtracting the frozen water content from the total water and was defined as the dilatometric bound water (BW_d) in this study.

A constant head piston press (Triton Electronics Ltd., type 147) was employed to find the bound water as defined by Lee [4]. About 350–400 ml samples of the original and the conditioned sludges were placed in the compression chamber. Hydraulic pressure of 1000 psi was then applied to the piston via the action of a gear pump. Owing to compression by the piston, filtrate was squeezed out and its weight automatically recorded by an electronic balance connected to a personal computer. After all the removable moisture had been removed, the residual moisture in the sludge cake was measured and defined as the expressed bound water (BW_e).

2.6. Flocs size distribution and fractal dimension analysis

The sizes and distributions of sludge flocs were measured using an image-analysis system (Galai ScanArray-2, Israel). The images of flocs were transferred to the analyzer, in which software was loaded. More than 500 flocs were analyzed in this work.

Effective floc density ($\rho_f - \rho_l$) can be determined using Eq. (2), which relates a fractal radius function R , derived from n primary particles of radius

R_0 , to a proportionality (C) and the primary particle dry density (ρ_p):

$$\rho_f - \rho_l = c \left(\frac{R}{R_0} \right)^{D-3} (\rho_p - \rho_l) \quad (2)$$

where ρ_f is the wet floc density of particle and ρ_l is the density of liquid. The fractal dimension (D) can be determined according to the slope of logarithmic correlation between diameter and effective density of the floc [7]. Different fractal dimension values can accurately describe aggregation properties under various chemical dosages. Moreover, their magnitude are related to aggregate morphologies.

3. Results and discussion

3.1. Polymer dosage and dewatering characteristics

Fig. 1 displays performance curve describing the relationship between polymer dosage and dewaterability. As the polymer dose was increased from 0 to 15 mg l⁻¹, a decrease in CST readings occurred, and a particular reduction trend is evident in the 5–15 mg l⁻¹ range. The CST reading increased when the dose was increased from 30 to 60 mg l⁻¹. This performance curve indicates the optimal dose is 15 mg l⁻¹, and that doses exceed-

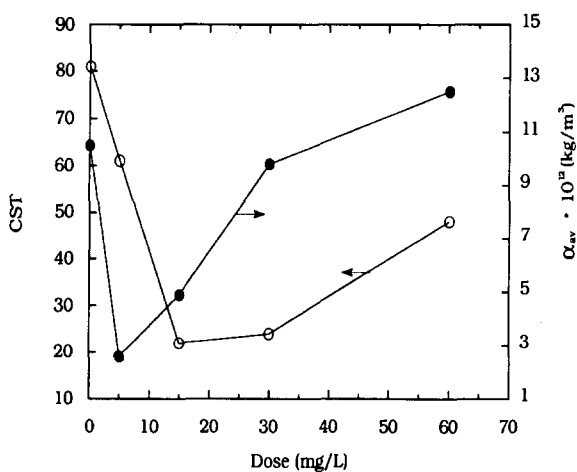


Fig. 1. Effects of cationic polymer (PC-320) dosage on CST and SRF of alum sludge.

ing 15 mg l⁻¹ can be regarded as the overdose condition according to CST measurements alone. In addition to CST testing, SRF has been evaluated and depicted in Fig. 1 as well. On the basis of the location of minimum α_{av} , the optimal dosage for dewaterability is 5 mg l⁻¹. This is apparently not consistent with that from CST test. Furthermore, dry solids content after centrifugation (CDS%), and the gravity settling dry solids content (GDS%) were evaluated and are listed in Table 2. Notably, the dosage of 5 mg l⁻¹ corresponds to the maximum values of GDS and CDS.

The data in Table 2 present the effects of cationic polymer dosage on the change of bound water content (BW_d and BW_e). Polymer conditioning significantly changes the bound water content of sludges. BW_d was reduced significantly when a polymer dosage of 5 mg l⁻¹ was used. However, BW_d immediately increased substantially when the polymer dosage exceeded 15 mg l⁻¹. Later though, when the dosage reached 60 mg l⁻¹, BW_d decreased slightly. In order to construct a conceptual description for interpreting the moisture change within a sludge, the terms “free water”, “interstitial water”, “surface water” and “internal water” as postulated by Tasng and Vesilind [15] are used in the present study. For circumstances in which the bound water content initially drops, Robinson and Knocke [3] suggested that bound water releases appear to correlate with the coagulation of sludge particles. Katsiris and Kouzeli-Katsirit [11] have proposed a hypothesis suggesting that removal of bound water following polymer application results from polymer replacing absorbed water molecules on particle surfaces. The moisture replacement capacity by 5 mg l⁻¹ polymer is unlikely to account for the marked reduction in BW_d . We therefore speculated that the action of inter-particles squeezing so as to induce the possible pore space collapse may attribute greatly to the reduction in BW_d . The increase in BW_d as the polymer dosage exceeded 15 mg l⁻¹ may be attributed to the increase of the physically bound moisture fraction (such as interstitial water) as the polyelectrolyte agglomerates fine particles into larger flocs [10]. Visual observation and dry solids content analysis of gravity settling (GDS) provide a clear support for this

Table 2
Effect of cationic polymer (PC-320) on the characteristics of sludge sample

Dose (mg l ⁻¹)	CST (s)	α_{av} ($\times 10^{12}$ m kg ⁻¹)	GDS (%)	CDS (%)	BW_d (kg-H ₂ O/kg-ds)	BW_e (kg-H ₂ O/kg-ds)
0	81	10.5	7.0	22.8	1.13	0.48
5	61	2.6	7.9	28.3	0.63	0.21
15	22	4.9	5.1	28.3	2.16	0.73
30	24	9.8	4.5	27.0	3.68	0.44
60	48	12.5	3.3	27.1	3.26	0.47

BW_d : Bound water content measured by dilatometer.

BW_e : Bound water content measured by expression test.

kg-ds: Weight of dry sludge.

GDS: Dry solids content after gravity settling.

CDS: Dry solids content after centrifugation.

explanation. Also, Table 2 reveals that conditioning at dosage under 30 and 60 mg l⁻¹ formed a cloud-like sludge suspension, thereby causing a low GDS value. As a result, we contend that the bound water depletion and interstitial water formation mechanisms coexist in the sludge conditioning system. Bound water content after sludge conditioning is the net change in moisture content, decreased by bound water depletion and increased by interstitial water formation.

Table 2 also shows the bound water data measured by expression test. Two things are noticeable. First, values of BW_e are smaller than BW_d , supporting the conclusion that bound water is an operationally defined value [4]. Second, there are similar patterns in the two tests for dependence of bound water content on polymer dose. This observation correlates with previous findings in which a certain sequence was shown to exist between bound water contents measured by different methods [16].

Fig. 2 presents a bound water sequence of polymer-conditioned sludge. The intrafloc water content (W_i) is defined as the value obtained by using the mass balance consideration while comparing the measured values of wet density and dry density of sludge. The three curves in Fig. 2 show that polymer conditioning significantly affects the water distribution in sludge. Theoretically, differences between W_i and BW_d should be equal to the amount of free water (or frozen water) that can be removed by mechanical dewatering equipment [15]. Differences between BW_d and BW_e are

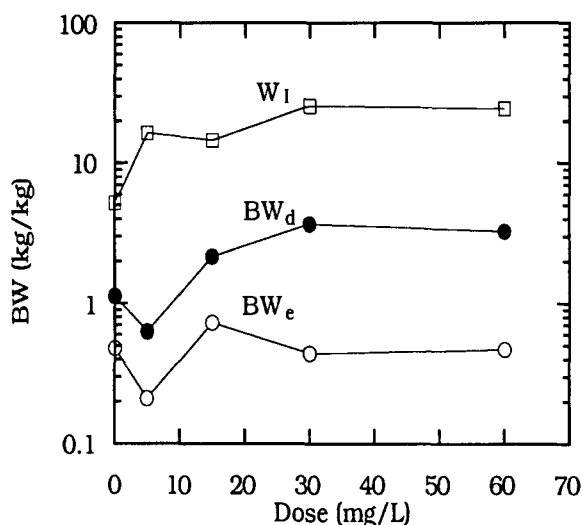


Fig. 2. Bound water sequence of original and conditioned sludge. Reported BW_d are an average of triplicate measurements. Reported BW_e are average of duplicate measurements.

approximately equal to the part of interstitial water trapped tightly within floc structures and on floc surfaces. This water remained unfrozen as low as -20°C but could still be removed by high centrifugal strain [17] or compression to 1000 psi in this study. Furthermore, residual water remaining in the sludge cake after compression is internal water, which can be regarded as the upper limit for the performance of any mechanical dewatering device [4].

Table 3 presents W_i and water removal ratios in response to the polymer dosage. With the water

Table 3

Effect of polymer dose on the intrafloc moisture content, moisture removal efficiency and physical properties of alum sludge

Dose (mg l ⁻¹)	W _i (kg kg ⁻¹)	R _f (%)	R _e (%)	FD value
0	5.23	78.4	90.8	1.18
5	16.40	96.2	98.7	1.04
15	14.64	85.2	95.1	1.39
30	25.54	85.6	98.3	1.39
60	24.50	86.7	98.1	1.83

$$R_f = ((W_i - BW_d)/W_i) \times 100 \quad R_e = ((W_i - BW_e)/W_i) \times 100$$

sequence in Fig. 2, the ratio (R_f) of the differences between W_i and BW_d divided by intrafloc water content can be calculated. Table 3 shows the ratio (R_e) of the differences between W_i and BW_e divided by W_i . The largest values of R_f and R_e both occurred at a dosage of 5 mg l⁻¹. This occurrence strongly supports the contention that cationic polymer changes the water distribution within alum sludge, thereby easing the release of free water.

On the basis of the bound water content, dewatered dry solid content and moisture removal efficiency, the 5 mg l⁻¹ polymer dosage is the most appropriate optimal dosage for dewatering sake. This is consistent with the optimal dosage determined via SRF, but not via CST. Because of this consistency, Knocke et al. [18] suggested that a significant decrease in floc water content gives a corresponding decrease in filtration resistance and enhances the dewatering of sludge. Obviously, determining the polymer dosage on the basis of CST measurements may lead to overdosing.

3.2. Polymer dosage and sludge physical properties

The data in Fig. 3 illustrate the typical correlation between floc diameter and effective density obtained from analysis of free settling. These results clearly indicate that floc density increases with decreasing floc diameter.

The effect of polymer dosage on the average particle diameter of a sludge sample measured by image analysis is shown in Fig. 4. The average diameter of the conditioned sludge floc significantly increases with the polymer dosage. Recall that, in Fig. 3, the floc density will decrease with

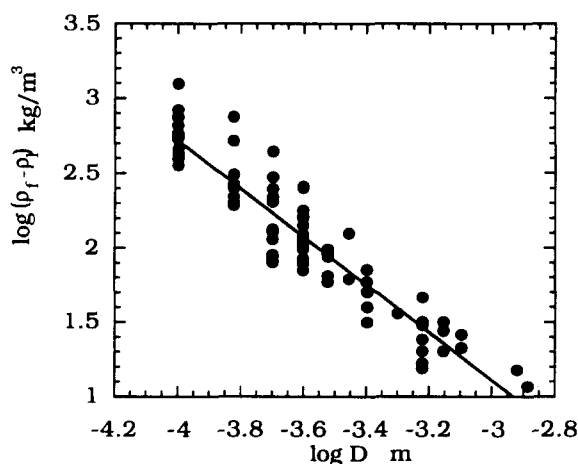


Fig. 3. Typical relationship between floc diameter (D) and effective density ($\rho_f - \rho_i$) at 30 mg l⁻¹ polymer dose.

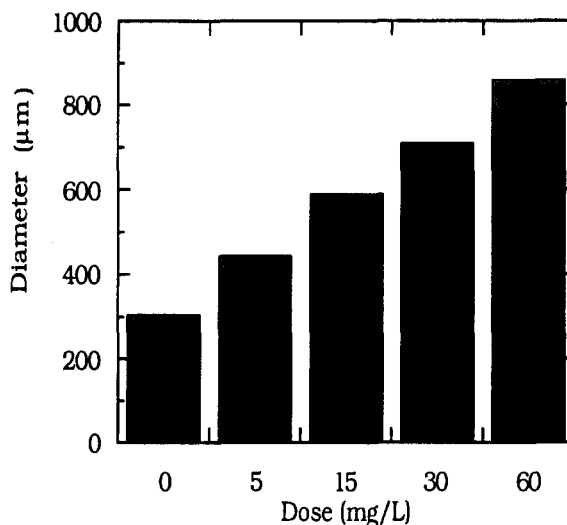


Fig. 4. Effects of polymer dosing on the average particle diameter of sludge measured by the image analysis.

increasing polymer dosage. Knocke and Kelley [19] have suggested that floc density significantly affects dewatering rates and high density sludge produces a better dewatering rate. Knocke et al. [15] have concluded that a significant decrease in floc water content would yield a corresponding improvement in specific filtration resistance. Therefore, based on our experimental results, we can infer that the poor dewatering performance (i.e. CST and SRF) and increased bound water

content are attributable to enlargement in floc size, and also to a decrease in floc density. However, when Knocke et al. [13] have used isopycnic centrifugation technology to measure the floc density of sludge; the results indicated that water released via polymer addition caused an increase in floc density. The discrepancy remaining between their results and ours may be due to the possible osmosis depletion of the intrafloc moisture.

Fractal dimension analysis is effective in investigating intrafloc structure. The theoretical values of fractal dimension vary from 1 to 3 and provide a useful index of the degree of floc compactness [20]. In Table 3, the fractal dimension values (FD value) of alum sludge range from 1.04 to 1.83 and seem to increase with polymer dosage. These results indicate that the degree of sludge floc compactness increases with polymer dosage. Interestingly, sludge flocs with higher polymer dosage have lower densities but more compact structures. Apart from the compactness evaluation, the fractal dimension is a measure of how the particle fills the space it occupies. For FD values close to 1 (such as the low dose case in this study), the type of aggregation process produced by primary particles forms lines similar to necklaces. In the high dose cases (in which FD is close to 2), the aggregation formation is similar to circular disks. Gill and Herrington [21] have provided a reasonable interpretation for the floc construction in the presence of polymer. At high polymer concentration, each floc is covered with several polymer molecules and, after the occurrence of many interparticle collisions, large granular flocs are produced as each particle is bound to several others. However, at lower polymer dosage perhaps, only one polymer molecular may be attached to each particle; thus, the particles join together like necklaces forming small but open flocs.

Finally, it can be concluded that the increase in bound water and the decrease in floc density are caused by variations in floc size and aggregation configurations, not the degree of floc compactness.

4. Conclusion

Results in this study indicate that using the CST measurement to determine the optimum polymer

dosage may lead to overdosing. This observation is confirmed by results obtained from vacuum filtration and bound water analysis of polymer-conditioned sludge. The magnitude of bound water content reflects the net change in moisture content decreased by water depletion and increased by water formation. Analysis of intrafloc water content, dialtometric bound water content and expressed bound water content, reveals that polymer conditioning significantly affects water distribution in sludge. According to the measured results of floc size and density, the poor dewatering performance and increasing bound water content are attributable to an enlargement in floc size, and also to a decrease in floc density. Moreover, the increase in the bound water and the decrease in floc density are caused by variations in both floc size and aggregation type, not the degree of floc compactness.

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